How functional traits, herbivory, and genetic diversity interact in Echinacea: implications for fragmented populations

PAMELA M. KITTELSON,^{1,5} STUART WAGENIUS,² REINA NIELSEN,¹ SANJIVE QAZI,¹ MICHAEL HOWE,¹ GRETEL KIEFER,³ AND RUTH G. SHAW⁴

¹Department of Biology and Environmental Studies, Gustavus Adolphus College, St. Peter, Minnesota 56082 USA ²Chicago Botanic Garden, Glencoe, Illinois 60022 USA ³Echinacea Project, Kensington, Minnesota 56343 USA

⁴Department of Ecology, Evolution and Behavior, University of Minnesota, St. Paul, Minnesota 55108 USA

Abstract. Habitat fragmentation produces small, spatially isolated populations that promote inbreeding. Remnant populations often contain inbred and outbred individuals, but it is unclear how inbreeding relative to outbreeding affects the expression of functional traits and biotic interactions such as herbivory. We measured a suite of 12 functional traits and herbivore damage on three genotypic cross types in the prairie forb, Echinacea angustifolia: inbred, and outbred crosses resulting from matings within and between remnant populations. Inbreeding significantly affected the expression of all 12 functional traits that influence resource capture. Inbred individuals had consistently lower photosynthetic rates, water use efficiencies, specific leaf areas, and had higher trichome numbers, percent C, and percent N than outbred individuals. However, herbivore damage did not differ significantly among the cross types and was not correlated with other leaf functional traits. Leaf architecture and low physiological rates of the inbred compared to outbred individuals imply poorer capture or use of resources. Inbred plants also had lower survival and fitness relative to outbred plants. Our results show that inbreeding, a phenomenon predicted and observed to occur in fragmented populations, influences key functional traits such as plant structure, physiology and elemental composition. Because of their likely role in fitness of individuals and ecological dynamics plant functional traits can serve as a bridge between evolution and community or ecosystem ecology.

Key words: Echinacea; fitness; functional traits; habitat fragmentation; herbivory; inbred; outbred; photosynthetic rate (A_{max}) ; remnant populations; specific leaf area (SLA); water use efficiency (WUE).

INTRODUCTION

The formation of small habitat islands during conversion of contiguous habitat tends to reduce the size of populations that remain (Saunders et al. 1991, Lienert and Fischer 2003). Isolation may continue to erode population size by decreasing gene flow and the probability of favorable demographic transitions (Benitez-Malvido 1998, Aguilar et al. 2006). These numerical effects can be compounded by genetic alterations. Small, isolated populations generally contain less genetic variation than larger interconnected populations (Honnay and Jacquemyn 2007) because random genetic drift and inbreeding are more likely (Aguilar et al. 2008); consequently increased homozygosity can result in inbreeding depression (Ellstrand and Elam 1993, Young et al. 1996). Depending on the strength of selection, population size, and distance, isolated populations also can be locally adapted (Nagy and Rice 1997, Etterson 2004, Leimu and Fischer 2008) and in these cases, reproduction between plants from distant remnants may

Manuscript received 2 September 2014; revised 11 November 2014; accepted 7 January 2015; final version received 28 January 2015. Corresponding Editor: N. J. B. Kraft.

result in outbreeding depression (Barrett and Kohn 1991, Waser and Price 1994).

Plant populations post-fragmentation likely comprise a mix of inbred and outbred plants. Despite the importance of understanding how populations respond to globally pervasive habitat fragmentation, few studies have measured how these genetically variable populations may respond demographically (Harrison and Bruna 1999, Lienert 2004, but see Wagenius et al. 2010). Even fewer have investigated variation in functional traits and interspecific interactions in relation to genetic differences due to inbreeding vs. outcrossing, and how these responses may be related to fitness (Kingsolver et al. 2001, Geber and Griffen 2003). Because habitat fragmentation can change patterns of reproduction, the prevalence of both inbred and outbred individuals in remnant populations can increase. Thus it becomes increasingly important to understand how genotypic classes express functional traits differently, and how these differences might influence the ecology and evolution of remaining populations. While many ecologists may take for granted that inbreeding depression results from compromised resource acquisition, little evidence exists for this.

⁵ E-mail: pkittels@gustavus.edu

Functional traits directly affect resource uptake and use, and can influence growth, survival, and reproduction (Geber 1990, Donovan and Ehleringer 1994). Functional traits can be physiological or morphological in nature and include photosynthetic rates, water use efficiency, photosynthetic nitrogen use efficiency, leaf area, and percent nitrogen. The expression of these traits can vary among genotypes by as much as 25-50% (Sultan and Bazzaz 1993, Sandquist and Ehleringer 2003). While physiologists have examined genotypic variability in functional traits, many studies are on agronomic or weedy species (Condon et al. 1993, Geber and Dawson 1997, Alonso-Blanco et al. 2009), annuals (Farris and Lechowicz 1990, Dudley 1996), or among plants growing in greenhouses; less is known about how functional traits vary within populations of perennial species and particularly how inbreds differ from outbreds in native, fragmented populations of perennial species (McGraw and Wulff 1983, Geber and Griffen 2003).

How inbreeding or outbreeding may differentially affect damage by insects also merits additional research. Inbreeding can increase susceptibility (Campbell et al. 2013) or resistance (Strauss and Karban 1994), or it can have no apparent effect (Nuñez-Farfan et al. 1996). Differential expression of functional traits also can mediate herbivory (Bello-Bedoy and Nuñez-Farfan 2011, Campbell et al. 2013). For example, foliar nitrogen or carbon may be strongly associated with herbivore susceptibility (Mattson 1980, Carmona et al. 2011). The thickness and hairiness of leaves also are strong adaptive defenses against herbivory (Agrawal and Fishbein 2006). Examining the dependence of herbivory on leaf functional traits and inbreeding can shed light on how plant populations respond to habitat fragmentation.

Isolated prairie remnants are a good system to evaluate how ecological and evolutionary processes influence native population persistence because they once covered vast regions of North America, but only 1– 4% now remain (Samson et al. 2004) and the biome is endangered (Hoekstra et al. 2005). Our research focuses on the perennial forb, *Echinacea angustifolia* (hereafter *Echinacea*). Inbreeding is likely to occur in isolated remnants (Wagenius 2000), and inbred *Echinacea* have 20% lower survival rate and ~66% lower fitness than non-inbred plants (Wagenius et al. 2010). However, it remains unknown whether functional traits or foliar herbivory differ among inbred and outbred *Echinacea*, and what relationship these traits might have to fitness.

Our research explores how leaf herbivory and functional traits vary among three genotypic classes of *Echinacea* plants differing in their degree of inbreeding. The three classes include plants derived from mating between plants from different prairie remnants (B), random mating between plants from the same prairie remnant (W), and sib-mating (I). Genotypic class W is likely typical of populations prior to fragmentation, while inbred plants (class I) may currently predominate in small, highly isolated remnants, and class B may become increasingly common because pollinators must travel farther to forage and because humans move seeds for restoration and other purposes. All three genotypic classes are now expected to be prevalent in our fragmented study area. The plants all grow in two experimental plots within a field undergoing restoration to a prairie community. Specifically, we address (1) How does a suite of functional traits differ in inbred and outbred plants, and how do these morpho-physiological traits relate to one another? (2) To what extent do plants from different cross types undergo different levels of foliar herbivory, and what interactions may exist between functional traits and leaf herbivory? We also discuss the relationship of these traits to existing measures of *Echinacea* fitness, which can yield insight into mechanistic processes that may underlie survival, reproduction, and adaptation (Ackerly et al. 2000, Caruso et al. 2005, Agrawal et al. 2008).

Methods

Study organism, cross types, and experimental plots

Narrow-leaved purple coneflower, Echinacea angustifolia (DC.), is an herbaceous perennial forb native to the Great Plains of North America. Conversion of prairie to agriculture has resulted in significant habitat loss for Echinacea. Small remnant populations that persist primarily in roadside ditches or areas inaccessible to farming can contain as few as three and up to hundreds of individuals (Wagenius 2000). Echinacea is selfincompatible and pollinated primarily by native solitary bees (Wagenius 2000, Wagenius and Lyon 2010). Populations may be isolated ≥ 1000 m from other remnant populations and successful pollination is significantly lower for spatially isolated Echinacea (Ison 2010, Wagenius and Lyon 2010). Fecundity is low in small populations often because compatible pollen donors are few (Wagenius 2004, 2006, Wagenius and Lyon 2010) and because flowering time is asynchronized (Ison 2010). Genetic differentiation exists within and among many Echinacea populations (Baskauf et al. 1994, Wagenius 2000, Kapteyn et al. 2002, Still et al. 2005, Ison 2010). Genetic diversity can be high in some western Minnesota prairie remnants suggesting little inbreeding in these sites (Ison 2010). Recruitment is limited only to seeds that gravity disperse. In early spring, plants sprout from a single taproot to form one to many rosettes. First flowering occurs at least three years after germination and then intermittently throughout an individual's lifetime. Plants generally make one inflorescence during the growing season. Seasonal senescence occurs in fall.

To assess how variation in mating affects plant traits and fitness, seedlings were generated from three types of crosses: (1) between maternal half sibs (inbred), (2) within a remnant, but not sharing a maternal or paternal parent (outbred), and (3) between individuals from different remnants (also outbred). Individuals from six remnant populations served as maternal parents, and were mated to multiple pollen donors from the same six remnants. We refer to individuals from these cross types as inbred (I), within remnant (W), and between remnants (B), respectively. Two experimental plots were established in prairie restoration on a former agricultural field in Douglas County, Minnesota (near 45°49' N, 95°42.5' W) from greenhouse grown seedlings in 2001 (INB1; N = 518) and 2006 (INB2; N = 1136). Following germination, seedlings of each cross type were randomly planted to positions 0.5 m apart and in rows 1 m apart. Management includes biyearly spring burns, mowing, periodic weeding of exotic invasive species and hand broadcasting of two native grasses. See Wagenius (2000), Wagenius et al. (2010), and Ridley et al. (2011) for more information on the species, crosses and study site.

Functional traits

We measured a suite of functional traits on Echinacea leaves growing in the two plots to address how inbred and outbred plants differed in functional trait expression. In late June/early July, and again in August 2013, we measured photosynthetic rates on an area basis (expressed as A_{max} ; µmol CO₂·m⁻²·s⁻¹) on the thirdmost basal leaf on the largest rosette using a Li-Cor 6400XT (Li-Cor, Lincoln, Nebraska, USA; n = 110 11year-old plants in INB1; n = 448 7-year-old plants in INB2). Chamber conditions were set to ambient CO₂ (400 μ L/L) and a saturating photon flux density (1600 μ mol quanta·m⁻²·s⁻¹). Air temperature tracked ambient. We took measurements from 10:30–15:00, when A_{max} peaked and plateaued in our diurnal response curves (data not shown). Leaves were allowed to equilibrate in the chamber and readings were recorded only after stability criteria were met for A_{max} , conductance, and transpiration (\sim 30–100 s). Instantaneous water use efficiency (WUE) was calculated by dividing A_{max} by transpiration/10000 (mol H₂0). The leaf was marked with a twist tie so that other functional trait measurements of the same leaf could be obtained.

Trichome density of marked leaves was determined in late July based on the better of two digital photographs taken on black cardstock. Each photograph was analyzed using ImageJ 64 (*available online*).⁶ Edge trichomes along a 2-cm transect were counted starting 1 cm from the tip. We averaged two separate counts of trichomes. If the difference between the two counts was greater than 15, both counts were made again. Edge trichomes positively correlated with trichomes on the leaf blade (Pearson's correlation, r = 0.80, df = 17, P < 0.001), so we used edge trichome data only.

After recording photosynthetic rates during peak flowering (12–15 Aug), we removed the marked leaf to calculate leaf dry matter content (LDMC; mg/g),

specific leaf area (SLA; m^2/g), and photosynthetic rate on a mass basis (A_{mass} ; µmol CO₂·g⁻¹·s⁻¹). After removal, leaves were saturated and weighed within 15– 60 min (following Wilson et al. [1999] and Garnier et al. [2001]). Each sample was immediately pressed. Leaves were dried to a constant dry mass (g) at 45°C. Average leaf area to the nearest 0.001 cm² was found by averaging two measurements from a calibrated Li-Cor 3100. Then, we dried leaves again and ground each sample for analysis of percent N, percent C, and C:N at the University of Nebraska Ecosystems Analysis Lab. Photosynthetic nitrogen use efficiency (PNUE) was calculated as µmol CO₂/mol N. The length of longest leaf (basal or cauline if no basal leaves present) was used as a relative estimate of overall plant size.

Interactions between herbivory and genetic diversity

Herbivore damage was measured on the same plants once in late July 2013 by rating leaf damage as none, 0%missing; minor, 1–25% of leaf removed or chewed; half, 25–75% of leaf missing or chewed; major, 75–95% of leaf missing; and >95%. Overall damage (%) was calculated by summing the midpoint for each leaf (0%, 12.5%, 50%, 85%, and 97.5%) and then finding the average for all the leaves on the plant. Aphid load also was recorded using the following categories for each leaf: 1 aphid, 2–10 aphids, 11–80 aphids, and >80 aphids.

Statistical analyses

Factor analysis was implemented in JMP (SAS Institute, Cary, North Carolina, USA) to group potentially correlated leaf traits as response variables so we could perform subsequent analyses on uncorrelated factors (Browne 2001). In factor analysis a principal component analysis is first done to reveal a correlation structure that explains the most variance in our 13 response variables (June A_{max} , August A_{max} , August Amass, June WUE, August WUE, percent N, percent C, C:N, LDMC, SLA, PNUE, average trichome count, and herbivore damage) by each successive principal component. Next, a scree plot was used to determine the minimum number of components required to account for the majority of the variation in the response variables. We then used an orthogonal Varimax method to rotate the principal components axes to better identify underlying factors that contribute to response variables. The rotation revealed clustering of correlated variables (factors) and the correlation of each response variable to its factor (factor loading). Traits were considered members of a factor when the factor loading was >0.7 and if that trait was associated with only that factor. Each factor was completely uncorrelated with other factors.

To evaluate the extent to which cross type (I, W, B) affected expression of the functional traits and herbivory, we performed ANCOVAs with the factors identified in the analysis previously described as response variables and the following predictors: experimental plot (INB1

⁶ http://rsbweb.nih.gov/ij/

1.492

position), maternai	family, cross typ	e (indred, bet	ween-, and within	-remnant cros	sses), longest l	ear length, and the	ar interactions.
Trait	Plot	Row (plot)	Position (plot)	Maternal population	Cross	Longest leaf	$Plot \times cross$	Cross × longest leaf
Factor 1 Factor 2	7.629 0.021	2.568 3.767	3.733 3.981	1.509	43.88 8.005	5027 152.0	4.720 3.469	0.341

1.819

TABLE 1. ANCOVA data from the three factors identified by factor analysis, which tested the effects of location (plot, row, position), maternal family, cross type (inbred, between-, and within-remnant crosses), longest leaf length, and their interactions.

Notes: F values are shown (df = 24, 520); boldface type denotes values that are significant at P < 0.05. Factor 1 includes June average photosynthetic rate on an area basis (A_{max}), August A_{max} , June instantaneous water use (WUE), August WUE; Factor 2 includes average photosynthetic rate on a mass basis (A_{max}) percent *C*, leaf dry matter content (LDMC), specific leaf area (SLA), and photosynthetic nitrogen use efficiency (PNUE); Factor 3 includes percent N and C:N ratio.

2.735†

 $† \hat{P} = 0.06.$

0 546

Factor 3

vs. INB2), row (plot), position (plot), maternal population, cross type, longest leaf length, and two-way interactions of cross \times leaf length and cross \times plot (using JMP). Interaction terms test whether effects of cross type on functional traits depended on longest leaf length and whether the effects of cross type on leaf traits depended on the experimental plot (environment and age of plants). We specified leaf length and position as continuous variables while the remainder were categorical variables. If a response variable was not identified as part of a factor (see Results), then we did a separate ANCOVA with that variable as the response using the same model described above. We also did a separate ANCOVA to investigate the effect of cross type on the length of the longest leaf by using length as a response variable.

6.172

1.216

Relationships between the leaf functional traits and herbivory were compared using partial correlation coefficients from the factor analysis. Herbivory also was analyzed using a logistic regression of herbivore damage (0 or 1) as a function of continuous leaf traits (using JMP).

RESULTS

Three factors, linear combinations of the 13 response variables, accounted for 64% of the variation. Factor 1, which accounted for 26% of the total variation, showed high loading for the physiological traits June A_{max} , August A_{max} , June WUE, August WUE (factor loadings = 0.87, 0.88, 0.79, and 0.72, respectively). Factor 2, which accounted for 23% of the total variation, showed high loading for A_{mass} , percent C, LDMC, SLA, and PNUE (factor loadings = 0.87, 0.68, -0.63, 0.87, and 0.82, respectively). Factor 3 comprised percent N and C:N (factor loading 0.98 and -0.93, respectively) and accounted for 16% of the variation. Herbivore damage and average trichome number did not group with any other variable to form a factor.

Cross type significantly affected factor 1 functional traits, as did longest leaf length, location (row, position, and plot) and the cross by plot interaction (Table 1). Inbred (I) *Echinacea* had the lowest photosynthetic rate (Fig. 1a) and WUE (Fig. 1b) followed by within remnant crosses (W). Between remnant crosses (B) were

highest, but in some cases W were not different statistically from either I or B cross types (differences among cross types P < 0.001; Fig. 1).

1.287

24.18

The factor 2 grouping, comprising leaf morphological and metabolic traits, also were affected by cross type, longest leaf length, location, and a cross by plot interaction (Table 1). Inbred plants expressed the lowest composite values for factor 2 relative to B or W cross types, regardless of plot (Table 2). PNUE, SLA, and A_{mass} were each ~10% lower in inbred crosses than B or W, while percent C, LDMC, were 1–4% higher in inbred plants.

The main effects on factor 3, percent N and C:N, were longest leaf and location (row; Table 1). Cross type affected factor 3 marginally (Table 1); inbred cross types grouped on their own with a mean (\pm SE) of 0.118 \pm 0.095 while composite values of both the B and W cross types overlapped strongly with one another (-0.149 \pm 0.012 and -0.131 \pm 0.078, respectively), but not with I. On average percent N was at about 1.5% higher for inbred plants compared to B and W.

Three distinguishable groups of *Echinacea* cross types appeared after plotting ellipsoids that captured 70% of the variation in the three factors, and after accounting for plant size and location (Fig. 2). The between remnant cross types were relatively distinct from the inbred while the within remnant cross types were intermediate in the bivariate and three-dimensional plots (Fig. 2).

As noted, an interaction between cross type and experimental plot (INB1 or 2) contributed to differences in the composite functional traits represented by factors 1 and 2. Inbred plants expressed much lower composite values of factor 1 and factor 2 traits in the INB2 experimental plot with 7-year-old plants, than in INB1 plot, which comprised older 12-year-old plants (Table 2). However, the relative ranking of cross types remained the same in each experiment; inbred crosses always expressed the lowest composite scores followed by within and between remnant crosses. The experimental plot did not explain differences among plants with regard to factor 3 (P > 0.05) or interact significantly with cross type (Table 1).



FIG. 1. (a) Average photosynthetic rates on an area basis $(A_{\text{max}}; \text{mean} \pm \text{SE})$ and (b) instantaneous water use efficiency (WUE; mean \pm SE) for three cross types of *Echinacea* angustifolia (inbred, between and within remnant cross types) in both experimental plots (INB1 and 2) early summer (27 Jun-12 Jul) and during peak flowering (12–15 Aug). Regardless of month or plot, between-remnant cross types always had the highest A_{max} and WUE, while inbred crosses had the lowest, and within-remnant cross types were intermediate.

Average trichome density did not correlate with other factors (Table 1) or the other functional traits (factor analysis partial correlation coefficients, P > 0.05). However, average trichome density did vary significantly as a function of cross; inbred plants had, on average, 1 more trichome per 2 cm than W, and 3 more than B.

Inbred and within remnant plants had more trichomes (4–7 more trichomes per 2 cm on average) in the older INB1 garden than INB2, and had more trichomes than between remnant crosses.

Overall foliar herbivore damage (percentage of the total) was low in 2013; over 36% of the plants had zero damage. The proportion of plants with damage to any leaf was 64%, 66%, and 62% for between, within, and inbred crosses, respectively. Aphid occurrence was exceptionally low compared to other years (Ridley et al. 2011, Shaw et al., in press); aphids were found on only five of 590 plants, and on plants derived from all three cross types. Leaf damage did not load onto any factor, and neither cross type nor length of longest leaf explained patterns in Echinacea mean herbivore damage (ANCOVA, $F_{24,520} = 1.12$, P = 0.318). The proportion of leaves without damage vs. those with damage also did not differ as a function of cross or any functional trait (logistic regression, P > 0.05 for cross type and each trait). Herbivore damage was not correlated with any functional trait, such as nitrogen, A_{max} , or trichomes (partial correlation coefficients, P > 0.05). Longest leaf length also was not significantly different among the three cross types (ANCOVA, $F_{21, 547} = 3.47$, P = 0.33).

DISCUSSION

We found that the three *Echinacea* cross types differed significantly in the expression of functional traits that influence resource capture and plant performance. Inbred plants had the lowest composite means for the three factors, and for specific traits such as A_{max} , WUE, and PNUE, while between remnant cross types had the highest composite and individual means (Figs. 1 and 2; Table 2). Thus, our measurements of functional traits showed no negative effect of crossing between remnants; instead, the effect on metabolic rates and functional traits was positive (Figs. 1 and 2). Within-remnant cross types tended to express intermediate traits (Fig. 2).

High A_{max} , WUE, and PNUE like those for betweenremnant crosses suggest better capture or use of resources, which has translated into greater growth or reproductive success of other species (Geber and Dawson 1990, Dawson and Ehleringer 1993). For example, lower photosynthetic rates in *Amaranthus*

TABLE 2. Composite least-square means $(\pm SE)$ for factor 1 and 2 for each cross type (B, between-remnant crosses; W, within-remnant crosses; and I, inbred) and experimental plot (INB1, INB2).

Cross type, plot	Factor 1	Factor 2
B, INB1 B, INB2 W, INB1 W, INB2 I, INB1 L INB2	$\begin{array}{c} 0.549 \pm 0.185 \\ 0.686 \pm 0.079 \\ 0.056 \pm 0.207 \\ -0.042 \pm 0.079 \\ -0.170 \pm 0.164 \\ 0.625 \pm 0.082 \end{array}$	$\begin{array}{c} 0.118 \pm 0.196 \\ 0.269 \pm 0.080 \\ -0.063 \pm 0.219 \\ 0.093 \pm 0.081 \\ -0.074 \pm 0.174 \\ 0.451 \pm 0.174 \end{array}$

Note: Factor 1 is primarily the physiological traits (June and August A_{max} and A_{mass} and WUE), while factor 2 is a composite of A_{mass} , percent C, LDMC, SLA, and PNUE.

mutants were directly related to lower survivorship and fertility than those with higher photosynthetic rates (Arntz et al. 2000). Greater reproductive success also was observed in high-WUE Xanthium strumarium than for plants with low WUE (Farris and Lechowicz 1990). Greater WUE was correlated with higher survivorship of Encelia farinosa during drought (Ehleringer 1993), and high WUE was selected for in dry treatments of Cakile edentula (Dudley 1996). Echinacea cross types B and W also appear to conserve water better while photosynthesizing at a relatively high rate (Fig. 1), which could lead to greater survivorship and reproduction, especially in the dry periods that often challenge tallgrass prairie perennials (Weaver and Albertson 1936, Clark et al. 2002). It is important to note that our study occurred in one growing season, which represents a fraction of the lifetime of our plants. The consistency of functional traits within individuals across years merits attention (Arntz and Delph 2001, Geber and Griffen 2003), particularly in plants such as Echinacea, where aboveground structures resprout each year from a taproot.

Echinacea leaf architecture was significantly related to physiological rates, which is a pattern found in other studies (Reich et al. 1997, Poorter and Garnier 1999). Inbred plants had smaller SLA, which on average also had lower mass- and area-based rates of photosynthesis, and PNUE. Inbred Echinacea also had high percent C and LDMC, which was negatively related to photosynthetic rates (A_{mass}) , possibly because the additional molecules that build a heavier leaf were non-photosynthetic (Wilson et al. 1999). The fact that the length of the longest leaves were not different among the cross types indicates that inbred and outbred plants may possess externally similar leaves but that the leaves differ in terms of quality (percent C, LDMC and SLA). Higher LDMC, like in the inbred Echinacea, can indicate resource limitation induced by abiotic factors or competition (Kittelson et al. 2008). Gonzalo-Turpin and Hazard (2009) found that patterns in SLA explained survivorship and reproduction of Festuca eskia along an elevation gradient, specifically higher SLA was correlated with higher values of fitness. Higher SLA and lower LDMC, like those measured in plants from between remnant crosses rather than inbred Echinacea, are also associated with higher growth rates (Navas and Moreau-Richard 2005, Thébault et al. 2011). While there were slight differences among cross types for some of the specific factor 2 traits, even small differences in leaf traits may affect subsequent development and fitness (Dawson and Ehleringer 1993, Dudley 1996). Thus, we might expect that lower expression of morphological and physiological traits associated with factor 2 contributes to the reduced fitness of our same inbred Echinacea individuals found by Wagenius et al. (2010).

Inbred *Echinacea* leaves had slightly more N and lower PNUE than leaves from non-inbred plants.

Among plants of the same size, inbred Echinacea may require slightly more N to maintain a similar photosynthetic rate and foliar biomass, and/or allocate foliar N into more non-photosynthetic constituents such as cell walls relative to non-inbred plants (Lambers and Poorter 1992, Hikosaka 2004). While efficient use of nitrogen can increase plant fitness (Hikosaka 2004), the small differences in foliar nitrogen among Echinacea cross types may be less biologically meaningful. However, our study found patterns in foliar N and C similar to those documented for the same INB1 plants by Ridley et al. (2011), which suggests that patterns in Echinacea leaf N and C are robust across years. Longterm differences in leaf N among different genotypes may influence processes such as litter decomposition, e.g., more inbred leaves with higher N could increase leaf decomposition rates (Cornwell et al. 2008, Kaproth et al. 2013) or higher C:N leaves of B and W remnant plants may result in slower rates, such that the nutrients remain available to promote plant growth the following spring (Hobbie 1992, Kaproth et al. 2013).

Cross type altered functional traits, which had the potential to influence species interactions, especially herbivory. However, in 2013, there was relatively little insect damage on Echinacea, and cross type did not explain the slight differences in foliar damage. We also did not find any relationship between leaf damage and functional traits, even traits often associated with plant susceptibility such as foliar percent N, trichomes, or LDMC (Pontes et al. 2007, Dalin et al. 2008). It is unlikely that 2013 herbivore damage strongly influenced measurements of Echinacea functional traits. Interestingly, variation in aphid infestation among Echinacea cross types has been found. In June of 2005, Ridley et al. (2011) found significantly more aphids on inbred individuals and between remnant individuals in INB1 Echinacea vs. within-remnant crosses, and aphid loads positively correlated to leaf N. While 2013 leaf damage did not depend on cross type or other foliar traits, and did not influence plant functional traits, there can be temporal variation in aphid loads both within a year (Ridley et al. 2011) and among years (Shaw et al., in press), and an accumulation of herbivory could affect Echinacea survival and reproduction over time. The interaction of functional traits and herbivores remains an open question that will need to be answered in a year when herbivores are more prevalent.

Functional traits in factors 1 and 2 and trichome number differed between the plots (cross by plot interactions), and along dimensions of the plots, as expected if functional traits are sensitive to environmental variation. Instantaneous gas exchange and leaf size can be influenced strongly by environmental variation (Geber and Griffen 2003, McKown et al. 2014). Regardless, inbred plants consistently had the lowest values even if the relative magnitude of the differences among the crosses differed between experimental plots (Table 2). Despite the influence of environment, cross



FIG. 2. Factor representation of the three *Echinacea* cross types comparing each factor in pairwise plots (a–c) and using a scatterplot depicting all three factors (d). Each cross type is depicted by a colored ellipsoid that captures 70% of the data. Factor 1 includes June A_{max} , August A_{max} , June WUE, August WUE; Factor 2 includes A_{mass} (average photosynthetic rate on a mass basis) percent C, leaf dry matter content (LDMC), specific leaf area (SLA), and photosynthetic nitrogen use efficiency (PNUE); Factor 3 includes percent N and C:N ratio.

type has a strong effect on plant functional traits, suggesting that selection on genetic variability in functional traits could occur (see Geber and Griffen 2003). The higher mean values for functional traits of inbred plants in the older INB1 compared to younger INB2 (Table 2) is consistent with selection during the course of the experiment having eliminated more of the plants with low values from the older cohort. Differences in ontogeny may also contribute to the trait differences. While it is rare that a single genotype will do well in all environments and variation in traits may change with environmental contexts (Sultan and Bazzaz 1993), our results show that, regardless of location, inbreeding impairs physiological traits, reducing the capture and assimilation of resources and resulting in different ways leaves are constructed, and these alterations in functional traits may translate into plants with lower fitness (Wagenius et al. 2010).

Taken together our results suggest that inbreeding strongly influences functional traits, and these measured differences have the potential to mediate fitness. For example, after seven years of growth, inbred plants in the INB1 plot had the lowest rate of survival, 20% and 35% lower than between and within cross types, respectively (Wagenius et al. 2010). We measured the survivors of these three cross types and fitness of our same group of inbred plants from INB1 was ~66% less than B and W cross types up to 2010 (Wagenius et al. 2010), and $\sim 60\%$ less at the end of 2012 (Shaw et al., in press). Other studies have found that functional traits were related to fitness correlates in annuals, e.g., populations of Polygonatum arenastrum flowered earlier when A_{max} was high (Geber 1990, Geber and Dawson 1997) and higher physiological rates in cocklebur was positively related to reproduction (Farris and Lechowicz 1990). Variation in functional traits also may indirectly influence fitness correlates (Dudley 1996, Arntz and Delph 2001). It is surprising that our measure of plant size, length of the longest leaf, did not differ among cross types in 2013, given that plant fitness has varied so strongly. On one hand, length of longest leaf is a coarse measure of size that could miss other key aspects, such as rosette or leaf number. On the other hand, perhaps Echinacea leaf length is not as clearly related to past and present fitness as the other functional traits we measured, such as leaf quality. For perennial species, it can take years to estimate fitness and Echinacea is no exception because there is little variation in survival and reproduction in a single year; the vast majority of plants survive and only a small proportion of plants reproduce. Our experiment is unique, however, because for INB1, we can relate functional traits to direct measures of fitness that were calculated from a nine year dataset, allowing us to reveal potential mechanisms that underlie the inbreeding depression observed in Echinacea. The positive relationship between functional traits and past fitness suggests that functional traits may be a good single-year predictor of survival and reproduction in subsequent years.

Little is known about how functional traits vary within populations, let alone whether patterns in reproduction and functional trait expression influence fitness in fragmented populations. Our combined results indicate that to the extent that habitat fragmentation increases the prevalence of inbred plants in populations, the traits of these individuals could reduce resource capture and assimilation, and have potential to mediate impaired survival and reproduction. In our study, physiological performance and persistence of inbred plants is lower than outbred progeny derived from between and within remnant crosses. Inbred individuals simply operate more poorly than outbred plants.

Severe habitat reduction can result in small plant populations, which can lead to inbreeding within remnants as well as outbreeding. Outbreeding occurs when pollinators move long distances among fragments and when people move plants and seed, as is increasingly common in our study area. It remains unclear what patterns might exist for other species common to remnant populations since few studies have examined the effect of in- and outbreeding on underlying functional traits. However, assuming our results are transferable to other long lived perennials, then inbred plants in fragmented populations might express a group of functional traits that can reduce persistence and potentially fitness, thereby exacerbating the numerical and genetic effects of habitat fragmentation on populations. Some restorations collect and use only locally sourced seed from few remnants to retain locally adapted genotypes (Broadhurst et al. 2008, Leimu and Fischer 2008) while bolstering population size, but this practice could result in inbred populations with compromised functional traits, especially if restored populations remain small. Other restorations use seeds from distant sources, much greater than the 6 km distance between our farthest sites of origin. While we found no adverse effect of outcrossing among remnants on functional traits, we cannot generalize that between remnant crosses at substantially greater distances would perform similarly. Regardless, our results suggest that restoration or active management needs to consider balancing local adaptation with avoidance of inbreeding.

Plant functional traits bridge evolution and community dynamics or ecosystem processes. This study shows that inbreeding, a phenomenon predicted and observed to occur in fragmented populations, influences key organismal traits, such as plant structure, physiology, and elemental composition. These traits are known to influence nutritional quality for herbivores, competitive interactions, and soil-microbe dynamics providing evidence that widespread changes in the genetic composition of remnant plant populations have potential to influence community and ecosystem dynamics.

ACKNOWLEDGMENTS

We thank the many Team *Echinacea* field assistants who helped to plant the plots and contribute to an annual database. We appreciate helpful discussions with J. Cavender-Bares, M. Kaproth, R. Drevonsky, and W. Pearse. We gratefully acknowledge support from the LiCor Biosciences Environmental Education Fund (LEEF), Gustavus Adolphus College First Year Research Experience (FYRE) Award and National Science Foundation Awards: 1052165, 1051791, and a Research Opportunity Award.

LITERATURE CITED

- Ackerly, D. D., et al. 2000. The evolution of plant ecophysiological traits. BioScience 50:979–995.
- Agrawal, A. A., A. C. Erwin, and S. C. Cook. 2008. Natural selection on and predicted responses of ecophysiological traits of swamp milkweed (*Asclepias incarnata*). Journal of Ecology 96:536–542.
- Agrawal, A. A., and M. Fishbein. 2006. Plant defense syndromes. Ecology 87(Supplement):S132–S149.
- Aguilar, R., L. Ashworth, L. Galetto, and M. A. Aizen. 2006. Plant reproductive susceptibility to habitat fragmentation: review and synthesis through a meta-analysis. Ecology Letters 9:968–980.
- Aguilar, R., M. Quesada, L. Ashworth, Y. Herrerias-Diego, and J. Lobo. 2008. Genetic consequences of habitat fragmentation in plant populations: susceptible signals in plant traits and methodological approaches. Molecular Ecology 17:5177–5188.
- Alonso-Blanco, C., G. M. Aarts, L. Bentsink, J. J. B. Keurentjes, M. Reymond, D. Vreugdenhil, and M. Koornneef. 2009. What has natural variation taught us about plant development, physiology and adaptation? Plant Cell 21: 1877–1896.

- Arntz, A. M., and L. F. Delph. 2001. Pattern and process: evidence for the evolution of photosynthetic traits in natural populations. Oecologia 127:455–467.
- Arntz, A. M., E. H. DeLucia, and N. Jordan. 2000. Fitness effects of a photosynthetic mutation across contrasting environments. Journal of Evolutionary Biology 13:792–803.
- Barrett, S. C. H., and J. R. Kohn. 1991. Genetic and evolutionary consequences of small population size in plants: implications for conservation. Pages 3–30 in D. A. Falk and K. E. Holsinger, editors. Genetics and conservation of rare plants. Oxford University Press, New York, New York, USA.
- Baskauf, C. J., D. E. McCauley, and W. G. Eickmeier. 1994. Genetic analysis of a rare and a widespread species of *Echinacea* (Asteraceae). Evolution 48:180–188.
- Bello-Bedoy, R., and J. Nuñez-Farfan. 2011. The effect of inbreeding on defence against multiple enemies in *Datura* stramonium. Journal of Evolutionary Biology 24:518–530.
- Benitez-Malvido, J. 1998. Impact of forest fragmentation on seedling abundance in a tropical rain forest. Conservation Biology 12:380–389.
- Broadhurst, L. M., A. Lowe, D. Coates, S. Cunningham, M. McDonald, P. Vesk, and C. Yates. 2008. Seed supply for broadscale restoration: maximizing evolutionary potential. Evolutionary Applications 1:587–597.
- Browne, M. W. 2001. An overview of analytic rotation in exploratory factor analysis. Multivariate Behavioral Research 36:111–150.
- Campbell, S. A., J. S. Thaler, and A. Kessler. 2013. Plant chemistry underlies herbivore-mediated inbreeding depression in nature. Ecology Letters 16:252–260.
- Carmona, D., M. J. Lajeunesse, and M. T. J. Johnson. 2011. Plant traits that predict resistance to herbivores. Functional Ecology 25:358–367.
- Caruso, C. M., H. Maherali, A. Mikulyuk, K. Carlson, and R. B. Jackson. 2005. Genetic variance and covariance for physiological traits in *Lobelia*: are there constraints on adaptive evolution? Evolution 59:826–837.
- Clark, J. S., E. C. Grimm, J. J. Donovan, S. C. Fritz, D. R. Engstrom, and J. E. Almendinger. 2002. Drought cycles and landscape responses to past aridity on prairies of the northern Great Plains, USA. Ecology 83:595–601.
- Condon, A. G., R. A. Richards, and G. D. Farquhar. 1993. Relationships between carbon isotope discrimination, water use efficiency and transpiration efficiency for dryland wheat. Australian Journal of Agricultural Research 44:1693–1711.
- Cornwell, W. K., J. H. Cornelissen, K. Amatangelo, E. Dorrepaal, V. T. Eviner, O. Godoy, and M. Westoby. 2008. Plant species traits are the predominant control on litter decomposition rates within biomes worldwide. Ecology Letters 11:1065–1071.
- Dalin, P., J. Ågren, C. Björkman, P. Huttunen, and K. Kärkkäinen. 2008. Leaf trichome formation and plant resistance to herbivory. Pages 89–105 in A. Schaller, editor. Induced plant resistance to herbivory. Springer, New York, New York, USA.
- Dawson, T. E., and J. R. Ehleringer. 1993. Gender-specific physiology, carbon isotope discrimination, and habitat distribution in boxelder, *Acer negundo*. Ecology 74:798–815.
- Donovan, L. A., and J. R. Ehleringer. 1994. Potential for selection on plants for water-use efficiency as estimated by carbon isotope discrimination. American Journal of Botany 81:927–935.
- Dudley, S. A. 1996. Differing selection on plant physiological traits in response to environmental water availability: a test of adaptive hypotheses. Evolution 50:92–102.
- Ehleringer, J. R. 1993. Variation in leaf carbon isotope discrimination in *Encelia farinosa*: implications for growth, competition, and drought survival. Oecologia 95:340–346.
- Ellstrand, N., and D. R. Elam. 1993. Population genetic consequences of small population size: implications for plant

conservation. Annual Review of Ecology and Systematics 24: 217–242.

- Etterson, J. R. 2004. Evolutionary potential of *Chamaecrista fasciculata* in relation to climate change. I. Clinal patterns of selection along an environmental gradient in the Great Plains. Evolution 58:1446–1456.
- Farris, M. A., and M. J. Lechowicz. 1990. Functional interactions among traits that determine reproductive success in a native annual plant. Ecology 7:548–557.
- Garnier, E., B. Shipley, C. Roumet, and G. Laurent. 2001. A standardized protocol for the determination of leaf area and leaf dry matter content. Functional Ecology 15:688–695.
- Geber, M. A. 1990. The cost of meristem limitation in *Polygonum arenastrum*: negative genetic correlations between fecundity and growth. Evolution 44:799–819.
- Geber, M. A., and T. E. Dawson. 1990. Genetic variation in and covariation between leaf gas exchange, morphology, and development in *Polygonum arenastrum*, an annual plant. Oecologia 85:153–158.
- Geber, M. A., and T. E. Dawson. 1997. Genetic variation in stomatal and biochemical limitations to photosynthesis in the annual plant, *Polygonum arenastrum*. Oecologia 109:535– 546.
- Geber, M. A., and L. R. Griffen. 2003. Inheritance and natural selection on functional traits. International Journal of Plant Sciences 164:S21–S42.
- Gonzalo-Turpin, H., and L. Hazard. 2009. Local adaptation occurs along altitudinal gradient despite the existence of gene flow in the alpine plant species *Festuca eskia*. Journal of Ecology 97:742–751.
- Harrison, S., and E. Bruna. 1999. Habitat fragmentation and large-scale conservation: what do we know for sure? Ecography 22:225–232.
- Hikosaka, K. 2004. Interspecific difference in the photosynthesis-nitrogen relationship: patterns, physiological causes, and ecological importance. Journal of Plant Research 117:481– 494.
- Hobbie, S. E. 1992. Effects of plant species on nutrient cycling. Trends in Ecology and Evolution 7:336–339.
- Hoekstra, J. M., T. M. Boucher, T. H. Ricketts, and C. Roberts. 2005. Confronting a biome crisis: global disparities of habitat loss and protection. Ecology Letters 8:23–29.
- Honnay, O., and H. Jacquemyn. 2007. Susceptibility of common and rare plant species to the genetic consequences of habitat fragmentation. Conservation Biology 21:823–831.
- Ison, J. L. 2010. Pollination of *Echinacea angustifolia*: effects of flowering phenology and spatial isolation. Dissertation. University of Illinois, Chicago, Illinois, USA.
- Kaproth, M. A., M. B. Eppinga, and J. Molofsky. 2013. Leaf litter variation influences invasion dynamics in the invasive wetland grass *Phalaris arundinacea*. Biological Invasions 15: 1819–1832.
- Kapteyn, J., P. Goldsbrough, and J. Simon. 2002. Genetic relationships and diversity of commercially relevant *Echinacea*. Theoretical and Applied Genetics 105:369–376.
- Kingsolver, J. G., H. E. Hoekstra, J. M. Hoekstra, D. Berrigan, S. N. Vignieri, C. E. Hill, A. Hoang, P. Gibert, and P. Beerli. 2001. The strength of phenotypic selection in natural populations. American Naturalist 157:245–261.
- Kittelson, P. M., J. L. Maron, and M. Marler. 2008. An invader differentially affects leaf physiology of two natives across a gradient in diversity. Ecology 89:1344–1351.
- Lambers, H., and H. Poorter. 1992. Inherent variation in growth rate between higher plants: a search for physiological causes and ecological consequences. Advances in Ecological Research 23:187–261.
- Leimu, R., and M. Fischer. 2008. A meta-analysis of local adaptation in plants. PLoS ONE 3:e4010.
- Lienert, J. 2004. Habitat fragmentation effects on fitness of plant populations—a review. Journal for Nature Conservation 12:53–72.

- Lienert, J., and M. Fischer. 2003. Habitat fragmentation affects the common wetland specialist *Primula farinose* in north-east Switzerland. Journal of Ecology 91:587–599.
- Mattson. 1980. Herbivory in relation to plant nitrogen content. Annual Review of Ecology and Systematics 11:119–161.
- McGraw, J. C., and R. Wulff. 1983. The study of plant growth: a link between the physiological ecology and population biology of plants. Journal of Theoretical Biology 103:21–28.
- McKown, A. D., R. D. Guy, J. Klápště, A. Geraldes, M. Friedmann, Q. C. B. Cronk, Y. A. El-Kassaby, S. D. Mansfield, and C. J. Douglas. 2014. Geographical and environmental gradients shape phenotypic trait variation and genetic structure in *Populus trichocarpa*. New Phytologist 201:1263–1276.
- Nagy, E. S., and K. J. Rice. 1997. Local adaptation in two subspecies of an annual plant: implications for migration and gene flow. Evolution 51:1079–1089.
- Navas, M. L., and J. Moreau-Richard. 2005. Can traits predict the competitive response of herbaceous Mediterranean species? Acta Oecologia 27:107–114.
- Nuñez-Farfan, J., R. A. Cabrales-Vargas, and R. Dirzo. 1996. Mating system consequences on resistance to herbivory and life history traits in *Datura stramonium*. American Journal of Botany 83:1041–1049.
- Pontes, L. D. S., J. F. Soussana, F. Louault, D. Andueza, and P. Carrere. 2007. Leaf traits affect the above-ground productivity and quality of pasture grasses. Functional Ecology 21:844–853.
- Poorter, H., and E. Garnier. 1999. Ecological significance of inherent variation in relative growth rate and its components. Pages 81–120 in F. I. Pugnaire and F. Valladares, editors. Handbook of functional plant ecology. Marcel Dekker, New York, New York, USA.
- Reich, P. B., M. B. Walters, and D. S. Ellsworth. 1997. From tropics to tundra: global convergence in plant functioning. Proceedings of the National Academy of Sciences USA 94: 13730–13734.
- Ridley, C., H. H. Hangelbroek, S. Wagenius, J. Stanton-Geddes, and R. Shaw. 2011. The effect of plant inbreeding and stoichiometry on interactions with herbivores in nature: *Echinacea angustifolia* and its specialist aphid. PLoS ONE 6: e24762.
- Samson, F. B., F. L. Knopf, and W. R. Ostlie. 2004. Great Plains ecosystems: past, present, and future. Wildlife Society Bulletin 32:6–15.
- Sandquist, D. R., and J. R. Ehleringer. 2003. Population- and family-level variation of brittlebush (*Encelia farinosa*, Asteraceae) pubescence: its relation to drought and implications for selection in variable environments. American Journal of Botany 90:1481–1486.

- Saunders, D. A., R. J. Hobbs, and C. R. Margules. 1991. Biological consequences of ecosystem fragmentation: a review. Conservation Biology 5:18–32.
- Shaw, R. G., S. Wagenius, and C. J. Geyer. *In press. Echinacea angustifolia* and its specialist aphid: the roles of plant phenotype and genotype. Journal of Ecology.
- Still, D. W., D.-H. Kim, and N. Aoyama. 2005. Genetic variation in *Echinacea angustifolia* along a climatic gradient. Annals of Botany 96:467–477.
- Strauss, S. Y., and R. Karban. 1994. The significance of outcrossing in an intimate plant–herbivore relationship. II. Does outcrossing pose a problem for thrips adapted to the host-plant clone? Evolution 48:465–476.
- Sultan, S. E., and F. A. Bazzaz. 1993. Phenotypic plasticity in *Polygonum persicaria*. I. Diversity and uniformity in genotypic norms of reaction to light. Evolution 47:1009–1031.
- Thébault, A., F. Gillet, H. Müller-Schärer, and A. Buttler. 2011. Polyploidy and invasion success: trait trade-offs in native and introduced cytotypes of two Asteraceae species. Plant Ecology 212:315–325.
- Wagenius, S. 2000. Performance of a prairie mating system in a fragmented habitat: self-incompatibility and limited pollen dispersal in *Echinacea angustifolia*. Dissertation. University of Minnesota, Minneapolis, Minnesota, USA.
- Wagenius, S. 2004. Style persistence, pollen limitation, and seed set in the common prairie plant *Echinacea angustifolia* (Asteraceae). International Journal of Plant Sciences 165: 595–603.
- Wagenius, S. 2006. Scale dependence of reproductive failure in fragmented *Echinacea* populations. Ecology 87:931–941.
- Wagenius, S., H. Hangelbroek, C. Ridley, and R. Shaw. 2010. Biparental inbreeding and interremnant mating in a perennial prairie plant: fitness consequences for progeny in their first eight years. Evolution 64:761–771.
- Wagenius, S., and S. P. Lyon. 2010. Reproduction of *Echinacea* angustifolia in fragmented prairie is pollen limited but not pollinator limited. Ecology 91:933–942.
- Waser, N. M., and M. V. Price. 1994. Crossing-distance effects in *Delphinium nelsonii*: outbreeding and inbreeding depression in progeny fitness. Evolution 48:842–852.
- Weaver, J. E., and F. W. Albertson. 1936. Effects on the great drought on the prairies of Iowa, Nebraska, and Kansas. Ecology 17:567–639.
- Wilson, P. J., K. Thompson, and J. G. Hodgson. 1999. Specific leaf area and leaf dry matter content as alternative predictors of plant strategies. New Phytologist 143:155–162.
- Young, A., T. Boyle, and A. H. D. Brown. 1996. The population genetic consequences of habitat fragmentation for plants. Trends in Ecology and Evolution 11:413–418.